

1 Response of soil aggregate disintegration to the different content of 2 organic carbon and its fractions during splash erosion

3 Abstract

4 Aggregate disintegration is a critical process in soil splash erosion. However,
5 the effect of soil organic carbon (SOC) and its fractions on soil aggregates
6 disintegration is still not clear. In this study, five soils with similar physical and
7 chemical properties and different contents of SOC have been used. The effects of
8 slaking and mechanical striking on splash erosion were distinguished by using
9 deionized water and 95% ethanol as raindrops. The simulated rainfall experiments
10 were carried out in four heights (0.5, 1.0, 1.5, and 2.0 m). The result indicated that
11 the soil aggregate stability increased with the increases of SOC and light fraction
12 organic carbon (LFOC). The relative slaking and the mechanical striking index
13 increased with the decreases of SOC and LFOC. The reduction of macroaggregates
14 in eroded soil gradually decreased with the increase of SOC and LFOC, especially in
15 alcohol test. The amount of macroaggregates ($>0.25\text{mm}$) in deionized water tests
16 were significantly less than that in alcohol tests under the same rainfall heights. The
17 contribution of slaking to splash erosion increased with the decrease of heavy
18 fractions organic carbon (HFOC). The contribution of mechanical striking was
19 dominant when the rainfall kinetic energy increased to a range of threshold between
20 $9\text{ J m}^{-2}\text{ mm}^{-1}$ and $12\text{ J m}^{-2}\text{ mm}^{-1}$. This study could provide the scientific basis for deeply
21 understanding the mechanism of soil aggregates disintegration and splash erosion.

22 1. Introduction

23 Splash erosion is an initial stage and an important component of interrill erosion
24 (Kinnell, 2005; Van Dijk, Meesters, & Bruijnzeel, 2002). The aggregate
25 fragmentation caused by raindrops striking is the first critical process in splash

erosion (Legout, Leguedois, Le Bissonnais, & Lssa, 2005; Shainberg, Levy, Rengasamy, & Frenkel, 1992). Many studies have showed that raindrops striking can damage the soil structure, disperse and transport soil particles, thereby reducing soil permeability. The amount of soil particles splashed increases with the speed at which the raindrop hits the ground (Moss & Green, 1987). Fu et al. (2017) found that there are significant exponential relations between the distance of splash and the size of raindrops. Xiao et al. (2017) reported that the contribution of slaking decreasing with the increased of the rainfall kinetic energy in splash, and that the contribution of mechanical striking was opposite. It was also found that the soil texture plays vital role in soil aggregates disintegration during splash erosion when the soil organic carbon (SOC) was low (Xiao et al., 2018).

Soil aggregate stability is a crucial physical indicator that determine disintegration resistance in soil erosion process, it determines the resistance of the soil to erosion (Barthes & Roose, 2002; Bronick & Lal, 2005; Nichols & Toro, 2011). The SOC plays an important role in formation of aggregates (Reeves, 1997; Smith & Petersen, 2000; Tisdall & Oades, 1982). Furthermore, the physical protection of aggregates is one of the main stabilizing mechanisms of SOC (Feller & Beare, 1997; Li & Pang, 2014; Tisdall & Oades, 1982). Emerson (1967) found that 90% of SOC in the topsoil was in aggregates. Six et al. (1998; 2004) reported that the SOC is one of the most important cementing materials in aggregates. Tisdall and Oades (1982) proposed the soil aggregate theoretical model, which states that microaggregate is an important precondition to form macroaggregate.

In order to distinguish the influence of SOC on the formation of aggregates, some researchers used physical or chemical methods to classify SOC into different components (Barrios, Buresh, & Sprent, 1996; Dhillon & Van Rees, 2017; Guan et al, 2018; Six, Paustian, Elliott, & Combrink, 2000; Xiang, Zhang, & Wen, 2015).

Density fractionation of SOC is a common way to separate the SOC into light fraction organic carbon (LFOC) and heavy fraction organic carbon (HFOC) (Christensen, 1992; Turchenek & Oades, 1979; Wagai, 2009). The LFOC was labile fractions that represent an intermediate organic carbon pool between humified organic matter and undecomposed residues (Janzen, Campbell, Brandt, Lafond, & Townley Smith, 1992). The HFOC was stable that has lower carbon concentrations and slow decomposition rate and transformation rate (Golchin, Clarke, Oades, & Skjemstad, 1995a; Golchin & Oades, 1995b; Hassink, 1995; John, Yamashita, Ludwig, & Flessa, 2005). Oades (1984) and Elliott (1986) found that the roots and the fungal hyphae in LFOC can promote the formation of aggregates directly. These studies focused on the formation process of aggregates or the influences of different components of organic carbon on the particle size of aggregates (Guan et al., 2018; Holeplass, Singh, & Lal, 2004; Xiang et al., 2015). However, the role of SOC and its fractions in inhibiting the destruction of aggregates during splash erosion is still not clear.

The purposes of this study were (i) to explore the influence of SOC and its fractions on the disintegration of aggregates; (ii) to quantify the effects of SOC on contribution of slaking and mechanical striking to splash erosion.

2. Materials and methods

2.1 Study area and sampling

Five soils were collected from Fuxian (36°03'~36°04'N, 109°08'~109°09'E) in Shaanxi province, China. The soils with distinct SOC contents, due to different years after conversion from cropland to forestry, were selected. The sampling areas belong to hilly and gully regions of the Loess Plateau and the main soil type is loessial soil. The average annual temperature and precipitation in this area are 9°C and 576 mm, respectively. The main types of land use were forest. The vegetation was mainly

78 *Ulmus pumila*, *Betula platyphlla*, *Prunus*, *Populus davidiana* and so on.

79 Soil samples were collected from the uppermost 20 cm layer, air-dried and
80 sieved through a 5 mm sieve to remove roots and gravels. The basic physical-
81 chemical properties of the soils in the experiment are showed in Table 1.

82 2.2 Experimental design

83 The rainfall device is composed of three parts, including rainfall liquid supply,
84 raindrop generation and supporting frame (Figure 1). The raindrop generation part
85 was a cylinder with diameter of 30 cm which is made from smooth steel. Thirty-nine
86 rainfall needles with a diameter of 0.6 mm were evenly distributed on the bottom of
87 the cylinder. The fall height was controlled by adjusting the height of the supporting
88 frame. Splash pan is an inverted cone device with an outer ring (30 cm diameter on
89 top, 10 cm diameter on bottom, and 30 cm in height) and an inner ring (10 cm
90 diameter and 10 cm in height) in the middle. Some small holes were drilled on the
91 bottom of inner ring for drainage. The outer ring and the inner ring were connected
92 to a smooth slope. An outlet was installed at the end of slope for water and sediment
93 collection. The detailed description of rainfall device can be found in Xiao et al.
94 (2017, 2018).

95 Before packing soil in the inner ring, the ring bottom was filled with some gravel
96 with diameters of 1-2 cm to ensure free drainage. Then a filter paper was covered
97 over the gravel and air-dried soil was packed over the paper. The packed soil was
98 about 1.5 cm thick with bulk density of 1.20 g cm^{-3} and initial soil water content of
99 5%. The duration of simulated rainfall was 10 mins with an intensity of 60 mm h^{-1} .
100 The kinetic energy was simulated at four different rainfall heights, 0.5 m, 1.0 m, 1.5
101 m and 2.0 m. Each treatment replicated twice. The splashed soil sticking on the wall
102 of outer ring and slope was washed out with an injector, and then collected at the
103 outlet. The collected sediment was dried and weighted. After test, the remained

104topsoil of 0.5 cm thick in the splash pan was collected to measure the water stable
105aggregates after air-dried.

106 Non-uniform swelling of soil minerals after wetting has limited damage to
107aggregates in rainfall conditions (Almajmaie, Hardie, Acuna, & Birch, 2017; Le
108Bissonnais, 1996). At the same time, the loessal parent material of the test soils was
109the least swelling mineral. Therefore, the destruction of soil aggregates by non-
110uniform swelling after soil mineral wetting was neglected in short duration of this
111study. The aggregates disintegration was mainly caused by slaking and mechanical
112striking (Xiao et al., 2017; 2018). The aggregates disintegration by 95% ethanol was
113resulted mainly from the mechanical striking because of the limited slaking effects
114of ethanol. Two raindrop materials (deionized water and ethanol), were used to
115distinguish the slaking and mechanical striking.

1162.3 Measurement

117 Soil particle size distribution, pH value, and soil organic carbon were analyzed
118by a pipette method (Liu, 1996), the Rex Electric Chemical PHS-3E precision
119acidity meter (Shanghai Precision Scientific Instrument Co., Ltd, China) and the
120potassium dichromate oxidation-external heating method, respectively (Liu, 1996).
121The content of CaCO_3 was determined by using a gas volume method (Dreimanis,
1221962; Zhao et al., 2016). The macroaggregate (>0.25 mm) measurement adopted the
123wet sieving method improved by Yoder (1936). The light and heavy fractions of
124organic carbon were separated in NaI solution with the density of 1.7 g cm^{-3} (Elliott,
1251991).

126 LB method was used to measure the aggregate stability under three treatments:
127fast wetting (FW), slow wetting (SW) and mechanical breakdown by slaking after
128pre-wetting (WS) (Le Bissonnais, 1996). The soil samples were air dried and 3-5
129mm aggregates were selected. The 3-5mm aggregates are dried in a 40°C oven for

13024 hours to ensure they are at the same matrix potential. Then the aggregates were
 131treated in three different treatments. FW: the aggregates of 5 g were immersed in 50
 132mL deionized water, and the water was absorbed by pipette after 10 minutes. SW:
 133the aggregates with 5 g were gently placed on the matrix potential of -0.3 kPa for 30
 134minutes to ensure that the aggregates were wetted completely. SW: the aggregates of
 1355 g were immersed in 50 mL ethanol (95% in mass), and the ethanol was absorbed
 136by pipette after 10 minutes. And then transferred the aggregates to a 250 mL flask
 137filled with 200 cm³ deionized water, corked and stirred up and down for 20 times,
 138and the water was absorbed by pipette after 30 minutes. Transfer the soil aggregates
 139from the above three treatments to a sieve (0.05 mm) already immersed in alcohol
 140(95% in mass) and shake up and down 20 times. The aggregate retained in the sieve
 141was baked for 48 hours in the oven at 40 °C. The dried aggregate was passed through
 142the dry sieve of 3, 2, 1, 0.5, 0.25, 0.1 and 0.05 mm, and then measured for their size.
 143Each treatment is repeated three times.

144 Aggregate stability is expressed in terms of mean weight diameter (MWD).

$$145 \quad \text{MWD} = \sum_{i=1}^n w_i x_i \quad (1)$$

146where w_i is the weight fraction of aggregates in size class i with an average diameter
 147 x_i .

1482.4 Data analysis

149 The relative slaking index (RSI) and the relative mechanical breakdown index
 150(RMI) were used to evaluate the sensitivity of aggregates to slaking and mechanical
 151breakdown effects (Zhang & Horn, 2001).

$$152 \quad \text{RSI} = \frac{\text{MWD}_{\text{SW}} - \text{MWD}_{\text{FW}}}{\text{MWD}_{\text{SW}}} \quad (2)$$

$$RMI = \frac{MWD_{SW} - MWD_{WS}}{MWD_{SW}} \quad (3)$$

where MWD_{FW} , MWD_{WS} , and MWD_{SW} are the mean weight diameter obtained by the FW, WS, and SW treatments, respectively (Le Bissonnais, 1996). The larger of RSI or RMI, the higher sensitivity of the aggregates to slaking or mechanical breakdown.

The splash erosion rate was the splashed-out soil mass from the test area per unit area per unit time, which can be calculated with Eq. (4):

$$D = \frac{S}{At} \quad (4)$$

where D is the splash erosion rate ($g\ m^{-2}\ min^{-1}$), S is the mass of the splashed material (g), A is the test area (m^2) and t is the duration of the rain (min).

The rainfall kinetic energy was calculated by referring to the formula in Xiao et al. (2017; 2018). Alcohol and deionized water have different rainfall kinetic energy due to their characteristics. The raindrop parameters and rainfall kinetic energy are shown in Table 2.

All statistical analyses were performed by using Excel 2010 and SPSS 19.0. Soil aggregate stability indexes were analyzed with a variance analysis (ANOVA), and the others with the Pearson correlation analysis (i.e., splash erosion rate, contribution of slaking and mechanical striking, etc.).

1713. Results

1723.1 Aggregate stability indexes

The aggregate stability indexes for the five soils are shown in Figure 2. The MWD_{FW} , MWD_{WS} and MWD_{SW} ranged from 0.612 to 2.389, from 1.202 to 3.262, and from 1.935 to 3.367, respectively. The MWD values increased with the increase

176 of SOC contents. The aggregate stability values increased in the order of MWD_{SW}
 177 $> MWD_{WS} > MWD_{FW}$ for the five soils. It showed that the effect of chemical
 178 dispersion (SW) was the weakest of aggregate breakdown mechanisms, whereas
 179 slaking (FW) had the most effect on aggregate breakdown. The RSI and RMI
 180 decreased from 0.698 to 0.293, and from 0.325 to 0.033 with the increase of SOC,
 181 respectively. The value of RSI was larger than that of RMI for the five soils.

182 [Figure 3](#) indicated that MWD_{FW} , MWD_{WS} and MWD_{SW} had significant positive
 183 correlations with SOC contents, but no significant positive correlation with LFOC
 184 and HFOC. They were negatively correlated with clay and the content of $CaCO_3$.
 185 MWD_{FW} , MWD_{WS} and MWD_{SW} had no significant positive relationships with the
 186 contents of free-form Fe, amorphous Fe, free-form Al and amorphous Al. RSI and
 187 RMI had significant negative correlations with the contents of SOC and LFOC,
 188 while they had no significant corrections with free-form Fe, amorphous Fe, free-
 189 form Al and amorphous Al contents.

190 3.2 Splash erosion rate

191 Splash erosion rate increased with the increase of rainfall kinetic energy for both
 192 deionized water and alcohol raindrops ([Figure 4](#)). The relationships for five soils
 193 could be described by power functions, and the coefficient of determination (R^2) was
 194 higher than 0.94 for both deionized water and ethanol tests ([Table 3](#)). The coefficient
 195 of power function can serve as an indicator of erosion severity with higher values
 196 reflecting higher soil erodibility.

197 The splash erosion rate has no significant negative correlations with SOC and
 198 HFOC in both deionized water and ethanol tests ([Table 4](#)). The negative and positive
 199 correlations were found between splash erosion rate and LFOC in deionized water
 200 and ethanol tests, respectively. An exception was that the negative correlation was
 201 found for ethanol tests in 1.5 m rainfall height. Meanwhile, compared with SOC and

202HFOC, LFOC had weaker relations with splash erosion rate in deionized water tests.

2033. 3 Macroaggregates

204 The macroaggregate ($>0.25\text{mm}$) contents remained in splash pan after rainfall in
 205 ethanol tests were more than that in deionized water tests (Figure 5). The
 206 macroaggregate contents for eroded soils were less than those of the parent soil for
 207 both deionized water and ethanol tests. However, the macroaggregate contents were
 208 increasingly closer to the parent soil with the increase of soil organic carbon
 209 contents, and the trend was more obvious in ethanol tests. The macroaggregate
 210 contents decreased with the increase of kinetic energy in both deionized water and
 211 ethanol tests, whereas the kinetic energy had no such significant effects in the
 212 ethanol tests except for soil sample IV.

213 The positive correlations were found between the macroaggregate contents and
 214 SOC, LFOC and HFOC in deionized water tests (Table 4). The correlations between
 215 macroaggregate and HFOC were weaker than those of SOC and LFOC. However,
 216 there were no significant correlations between them in ethanol tests.

2173.4 Effects of slaking and mechanical striking on splash erosion

218 Figure 6 showed that the contribution rate of slaking and mechanical striking
 219 decreased from 75% to 25% and increased from 25% to 75% with the increase of
 220 rainfall kinetic energy, respectively. Meanwhile, when the rainfall kinetic energy was
 221 less than the range of critical values (between $9\text{ J m}^{-2}\text{ mm}^{-1}$ and $12\text{ J m}^{-2}\text{ mm}^{-1}$), the
 222 contribution of slaking has dominant impact on aggregates disintegration. When the
 223 rainfall kinetic energy was greater than the range of critical values, the contribution
 224 of mechanical striking to splash erosion is gradually greater than that of slaking.

225 Table 4 indicated that the contribution rate of slaking had negative correlations
 226 with SOC contents when the kinetic energy increased from 3 to $12\text{ J m}^{-2}\text{ mm}^{-1}$, and it
 227 had positive correlations when the kinetic energy changed between 15 to 18 J m^{-2}

228mm⁻¹. With the increase of the kinetic energy, the correlation coefficient decreased
 229from -0.774 to 0.061. There were no significant correlations between the
 230contribution rate of slaking and LFOC, meanwhile the correlation coefficient had an
 231increasing trend with the increase of the kinetic energy. The correlations between
 232contribution rate of slaking and HFOC was significantly negative when the kinetic
 233energy increased from 3 to 6 J m⁻² mm⁻¹. The correlation coefficient decreased from
 2340.900 to 0.671 with the increase of kinetic energy.

2354. Discussion

236 Generally, soil clay, SOC, CaCO₃, and Fe/Al oxides act as cementing agents
 237that affect the formation and stability of aggregates (An, Darboux, & Cheng, 2013;
 238Dimoyiannis, 2012; Le Bissonnais., 1996; Le Bissonnais & Arrouays, 1997). The
 239aggregate stability had significant positive correlation with SOC but not with the
 240contents of clay, CaCO₃, and Fe/Al oxides (Figure 3). In this study, SOC acted as the
 241main factor affecting the aggregate stability because the test soils had the similar
 242contents of clay, CaCO₃, and Fe/Al oxides (Table 1). The SOC and LFOC had
 243significantly negative correlations with RSI and RMI, illustrating that the sensitivity
 244of slaking and mechanical striking decreased with increases of SOC and LFOC.
 245Therefore, SOC, especially the LFOC, played an important role in resisting
 246disintegration of aggregates. The formation of soil aggregate relies on organic
 247materials, and the organic binding agents were mainly polysaccharides, roots and
 248fungal hyphae, strongly sorbed natural polymers, and so on (Sdall & Oades, 1982).
 249The roots and fungal in composition of LFOC could promote the formation of soil
 250aggregate directly (Elliott, 1986; Oades, 1984). Thus, the aggregate stability and the
 251organic binding agents increased with the content of LFOC.

252 The power function relationships between splash erosion rate and rainfall kinetic
 253energy is consistent with the conclusions of the previous researchers (Hu, Zhen, &

254 Bian, 2016; Sharma, Gupta, & Rawls, 1991; Xiao et al., 2017; 2018). There was a
 255 tendency that splash erosion rate was negatively correlated with SOC, LFOC and
 256 HFOC for both deionized water and ethanol tests. However, the correlations were
 257 not statistically significant. This could be caused by the relatively narrow ranges of
 258 SOC, LFOC and HFOC used in this study, or splash erosion might be not as
 259 sensitive to SOC and LFOC as the aggregate sensitivity to slaking and mechanical
 260 breakdown effects.

261 Raindrops hit the soil surface with a certain kinetic energy, which is often
 262 sufficient to breakdown soil aggregates and compact the soil surface (Moss & Green,
 263 1987). The deionized water raindrops had both slaking and mechanical striking
 264 effects on aggregate disintegration, whereas alcohol only had mechanical striking
 265 effects (Le Bissonnais, 1996). On the other hand, the kinetic energy of deionized
 266 water raindrops was greater than that of ethanol raindrops at the same fall height
 267 (Table 2). These results lead to the destructive capacity of deionized water raindrops
 268 were greater than that of ethanol.

269 The mechanical striking of raindrops on soil could be greatly reduced by
 270 vegetation cover (Lal, 1976; Adekalu, Olorunfemi, & Osunbitan, 2007; Kukal &
 271 Sarkar, 2010). However, vegetation could promote the accumulation of SOC,
 272 especially the LFOC in short term (Boone, 1994; Garcia, Hemanderz, Roldan, &
 273 Martin, 2002; Gil-Sotres, Trasar-Cepeda, Leiros, & Seoane, 2005). That resulted in
 274 the increase of aggregate stability (Figure 2), and counteracted the increase of
 275 slaking contribution. Finally, vegetation coverage could improve soil antierodibility
 276 by reducing both slaking and mechanical striking effects of raindrops.

277 There are some limitations for testing five soils developed from only one parent.
 278 The effects of SOC on soil aggregates disintegration may be different for different
 279 soil types due to interactive effects of other factors. Furthermore, the coupled effects

of other factors in aggregate breakdown during splash erosion also need to be researched in the future.

25. Conclusions

In this study, the simulated rainfall experiments for five soils with different SOC were carried out. The results indicated that the content of SOC and LFOC had substantial effects on aggregate stability. The RSI and RMI decreased as SOC and LFOC increased. The amount of macroaggregates in deionized water tests were significantly less than that in alcohol tests under the same rainfall heights. The reduction of macroaggregates in eroded soil gradually decreased with the increase of SOC and LFOC, especially in alcohol test. As the rainfall kinetic energy increased, the contribution of slaking to soil splash decreased while the contribution of mechanical striking increased. The range of critical values between $9 \text{ J m}^{-2} \text{ mm}^{-1}$ and $12 \text{ J m}^{-2} \text{ mm}^{-1}$ were found to determine the dominated contribution of slaking and mechanical striking to splash erosion.

Data sharing: Research data are not shared.

Conflict of interest: none

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Soil samples	Year of vegetation restoration / a	Longitude and latitude	Clay / %	Silt / %	Sand /%	CaCO ₃ / g kg ⁻¹	Soil moisture content/ %	Soil organic carbon/ g kg ⁻¹	Light fractions of soil organic carbon/ g kg ⁻¹	Heavy fractions of soil organic carbon/ g kg ⁻¹	pH (1:2.5)	Free-form Fe/g kg ⁻¹	Amorphous Fe/g kg ⁻¹	Free-form Al/g kg ⁻¹	Amorphous Al/g kg ⁻¹
I	3	N36°03.61', E109°08.99'	22.36	21.16	56.48	81.09	2.46	7.54	5.89	1.68	8.51	5.10	0.48	1.57	0.58
II	20	N36°03.60', E109°09.03'	23.54	25.23	51.23	82.53	1.18	13.34	10.45	2.86	8.36	7.91	0.45	0.73	1.17
III	100	N36°04.94', E109°08.71'	20.18	29.63	50.20	77.63	0.72	14.85	9.83	4.99	8.42	8.93	0.64	0.56	1.31
IV	80	N36°03.68', E109°08.87'	21.05	18.06	60.88	76.76	1.30	18.39	15.70	2.66	8.25	6.16	0.81	1.87	1.63
V	55	N36°03.85', E109°08.78'	22.12	15.84	62.04	79.98	0.92	21.69	17.88	3.78	8.37	4.66	0.39	1.70	0.44

474

Tab. 2 Rainfall kinetic energy for different fall heights

Liquid	Fall height/m	Time for 10 raindrops/s	Weight of 10 raindrops/g	Mean raindrop diameter/mm	Rainfall kinetic energy/J m ⁻² mm ⁻¹
Deionized water	0.5	9.91	0.09	2.62	1.77
	1.0	9.91	0.09	2.63	6.15
	1.5	9.91	0.09	2.62	12.06
	2.0	9.92	0.09	2.63	18.75
Ethanol	0.5	6.45	0.04	2.03	1.49
	1.0	6.23	0.04	2.03	5.00
	1.5	6.34	0.04	2.03	8.90
	2.0	6.29	0.04	2.03	13.06

475 Tab. 3 Nonlinear regression between rainfall kinetic energy and splash erosion rate in simulated rainfall tests by using deionized water and
 476 ethanol as raindrop

Number of soil samples	Deionized water raindrop	R ²	Ethanol raindrop	R ²
I	$D = 1.1959 e^{1.15}$	0.97	$D = 0.3055 e^{1.38}$	0.98
II	$D = 2.2951 e^{0.89}$	0.99	$D = 0.5391 e^{1.27}$	0.99
III	$D = 0.7207 e^{1.18}$	0.99	$D = 0.2026 e^{1.51}$	0.98
IV	$D = 1.8037 e^{1.03}$	0.97	$D = 0.5753 e^{1.19}$	0.94
V	$D = 0.6303 e^{1.26}$	0.99	$D = 0.2111 e^{1.41}$	0.96

Notes: D is the soil splash rate ($\text{g m}^{-2} \text{min}^{-1}$) and e is the rainfall kinetic energy ($\text{J m}^{-2} \text{mm}^{-1}$)

477 Tab. 4 Pearson correlation coefficients between soil organic carbon and its fractions related to the splash erosion rate, contribution of slaking and
478 macroaggregates (>0.25mm) contents

	Splash erosion rate								Contribution of slaking								Macroaggregates (>0.25mm) contents							
	Deionized water				Ethanol												Deionized water				Ethanol			
	D _{0.5}	D _{1.0}	D _{1.5}	D _{2.0}	D _{0.5}	D _{1.0}	D _{1.5}	D _{2.0}	C ₃	C ₆	C ₉	C ₁₂	C ₁₅	C ₁₈	W _{0.5}	W _{1.0}	W _{1.5}	W _{2.0}	W _{0.5}	W _{1.0}	W _{1.5}	W _{2.0}		
SOC	-0.251	-0.270	-0.310	-0.513	-0.036	-0.031	-0.502	-0.026	-0.774	-0.334	-0.126	-0.005	0.016	0.061	0.632	0.429	0.616	0.656	0.083	0.036	-0.144	-0.016		
LFOC	-0.158	-0.086	-0.131	-0.363	0.095	0.056	-0.490	0.120	-0.623	-0.129	0.072	0.188	0.201	0.245	0.616	0.431	0.606	0.619	0.177	-0.129	-0.072	0.050		
HFOC	-0.468	-.0822	-0.818	-0.791	-0.525	-0.353	-0.263	-0.576	-0.900*	-0.918*	-0.802	-0.733	-0.692	-0.671	0.316	0.163	0.289	0.408	-0.328	-0.349	-0.343	-0.267		

Note: D_{0.5}, D_{1.0}, D_{1.5}, D_{2.0} is splash erosion rate at different fall heights (0.5m, 1.0m, 1.5m and 2.0m), respectively; C₃, C₆, C₉, C₁₂, C₁₅ and C₁₈ is the contribution of slaking in different rainfall kinetic energy (3, 6, 9, 12, 15 and 18 J m⁻² mm⁻¹), respectively. SOC, LFOC and HFOC is soil organic carbon, light fractions organic carbon and heavy fractions organic carbon content, respectively. W_{0.5}, W_{1.0}, W_{1.5} and W_{2.0} is the contents of >0.25mm water stable aggregates at different heights (0.5m, 1.0m, 1.5m and 2.0m).

* Significant at 0.05 level of probability.

479 **Figure captions**

480 Figure 1. Schematic representation of the experiment device

481 Figure 2. Aggregates water-stability of five loessial soils developed from same parent

482 material with different soil organic carbon (SOC) contents. The SOC contents of

483 I, II, III, IV, and V is 7.54, 13.34, 14.85, 18.39, 21.69 g kg⁻¹, respectively

484 Different letters in the same set of data of the same color indicate significant

485 differences at 5% level.

486 Figure 3. Heatmap for the relationships between soil aggregate stability indexes and

487 soil properties. MW_{FW} and MW_{SW} denote the mean weight

488 diameters obtained after the fast-wetting (FW), pre-wetting and stirring (WS)

489 and slow wetting (SW), respectively; RSI and RMI denote relative slaking index

490 and relative mechanical breakdown index, respectively; SOC, LFOC and HFOC

491 denote soil organic carbon, light fractions organic carbon and heavy fractions

492 organic carbon content, respectively.

493 Figure 4. Relationships between rainfall kinetic energy of two kind of raindrops (A is

494 deionized water; B is ethanol) and splash erosion rate. I, II, III, IV and V were

495 five different tested soils, which were developed from the same

496 similar physicochemical properties and different organic carbon content because

497 of different vegetation restoration time.

498 Figure 5. The contents of macroaggregates (>0.25mm) in parent soil and

499 remained in splash pan after rainfall at different height (0.5, 1.0, 1.5, and 2.0 m)

500 with different raindrops (A is deionized water; B is ethanol). Different letters in

501 the same group of each soil sample indicate significant differences at the 0.05

502 level.

503 Figure 6. Changes of contribution rate of slaking and mechanical striking (S is the
504 contribution rate of slaking; M is the contribution rate of mechanical striking) to
505 splash erosion with rainfall kinetic energy. I, II, III, IV and V were five different
506 tested soils, which were developed from the s
507 physicochemical properties and different organic carbon conten
508 different vegetation restoration time.